

## ENGINEERING AND GINNING

### First Stage Mote System Particulate Emission Factors for Cotton Gins: Particle Size Distribution Characteristics

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#### ABSTRACT

**This report is part of a project to characterize cotton gin emissions from the standpoint of total particulate stack sampling and particle size analyses. In 2006 and again in 2013, the United States (U.S.) Environmental Protection Agency (EPA) published a more stringent National Ambient Air Quality Standard for particulate matter with nominal diameter less than or equal to 2.5  $\mu\text{m}$  ( $\text{PM}_{2.5}$ ). This created an urgent need to collect additional cotton gin emissions data to address current regulatory issues, because EPA AP-42 cotton gin  $\text{PM}_{2.5}$  emission factors were limited. In addition, current EPA AP-42 included combined mote cleaner  $\text{PM}_{10}$  (particulate matter with nominal diameter less than or equal to 10  $\mu\text{m}$ ) and total particulate emission factors and not individual mote system emission factors. The objective of this study was to characterize particulate emissions for 1<sup>st</sup> stage mote systems from cotton gins across the cotton belt based on particle size distribution analysis of total particulate samples from EPA-approved stack sampling methods. Average measured  $\text{PM}_{2.5}$ ,  $\text{PM}_6$ , and  $\text{PM}_{10}$  emission factors based on the mass and particle size analyses of EPA Method 17 total particulate filter and wash samples from five gins (14 total test runs) were 0.00063 kg/227-kg bale (0.0014 lb/500-lb bale), 0.0054 kg/bale (0.012 lb/bale), and 0.0091 kg/bale (0.020 lb/bale), respectively. The 1<sup>st</sup> stage mote system particle size distributions were characterized**

**by an average mass median diameter of 16.4  $\mu\text{m}$  (aerodynamic equivalent diameter). Based on system average emission factors, the ratio of  $\text{PM}_{2.5}$  to total particulate was 2.49%,  $\text{PM}_6$  to total was 21.6%, and  $\text{PM}_{10}$  to total was 36.0%.**

In 2006 and again in 2013, the United States (U.S.) Environmental Protection Agency (EPA) published a more stringent standard for particulate matter (PM) with a particle diameter less than or equal to a nominal 2.5- $\mu\text{m}$  ( $\text{PM}_{2.5}$ ) aerodynamic equivalent diameter (AED) (CFR, 2013). The cotton industry's primary concern with this standard was the limited cotton gin  $\text{PM}_{2.5}$  emissions data published in the literature and in EPA's AP-42, Compilation of Air Pollutant Emission Factors (EPA, 1996b). AP-42 was first circulated in 1972 and the last complete document revision was in 1995. Since 1995, only updates and supplements have been added. AP-42 contains air pollutant emission factors for more than 200 industrial sources of air pollution along with information on the processes conducted at these sources.

An emission factor is a relationship between a process and the amount of air pollution emitted by that process into the atmosphere (EPA, 1996b). Emission factors are usually defined as the weight of pollutant emitted per unit weight, volume, distance, or duration of the activity producing the pollutant (e.g., kilogram of particulate emitted per cotton bale ginned). These relationships have been established from source test data, modeling, material balance studies, and engineering estimates and are usually averages of all data that have been gathered for a particular process (EPA, 1996a).

EPA's AP-42 was developed to include emission factors for all criteria pollutants and additional pollutants beyond the scope of the National Ambient Air Quality Standards (NAAQS), including total PM,  $\text{PM}_{10}$  (PM with a particle diameter less than or equal to a nominal 10- $\mu\text{m}$  AED), and  $\text{PM}_{2.5}$ . Current AP-42 cotton gin emission factors are located in section 9.7 (EPA, 1996b). Further, Appendix B.1 of AP-42 contains particle size distribution (PSD)

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data and emission factors based on these PSDs (EPA, 1996c). The only PM<sub>2.5</sub> emission factors in the current AP-42 were listed in Appendix B.1 and were based on PSDs. The 1996 AP-42 version only contained cotton ginning PSD data for the battery condenser and combined lint cleaning systems. The information for the battery condenser system equipped with cyclones was based on two tests and the PSD data was determined using a UW Mark 3 Impactor. The information for the combined lint cleaning system equipped with cyclones was based on four tests. The total particulate concentration data was determined using EPA Method 5 and the PSD data was determined by using a Coulter Counter to process the Method 5 samples (Hughes et al., 1982). Hughes et al. (1982) did not specifically state whether the PSD results were based on both the Method 5 wash and filter samples, wash only, or filter only. Table 1 provides examples of the types of data that were provided in EPA’s AP-42 Appendix B.1.

Emission factors from EPA AP-42 developed prior to 2013 were assigned ratings to assess the quality of the data being referenced. The ratings ranged from A (excellent) to E (poor). The PSD data quality rating in the 1996 AP-42 for both the battery condenser and combined lint cleaning systems was E (EPA, 1996c).

Cotton ginners’ associations across the U.S. cotton belt, including the National, Texas, Southern, Southeastern, and California associations, agreed that there was an urgent need to collect additional PSD data on PM being emitted from cotton ginning system exhausts. Working with cotton ginning associations across the country, state and federal regulatory agencies, Oklahoma State University, and USDA-Agricultural Research Service (ARS) researchers developed a proposal and sampling plan that was initiated in 2008 to address this need. Buser et al. (2012) provided the details of this sampling plan. This article is part of a series that details cotton gin emission factors developed from coupling total particulate stack sampling concentrations and particle size analyses. Each manuscript in the series

addresses a specific cotton ginning system. The systems covered in the series include: unloading, 1<sup>st</sup> stage seed-cotton cleaning, 2<sup>nd</sup> stage seed-cotton cleaning, 3<sup>rd</sup> stage seed-cotton cleaning, overflow, 1<sup>st</sup> stage lint cleaning, 2<sup>nd</sup> stage lint cleaning, combined lint cleaning, cyclone robber, 1<sup>st</sup> stage mote, 2<sup>nd</sup> stage mote, combined mote, mote cyclone robber, mote cleaner, mote trash, battery condenser, and master trash. This manuscript reports on the characterization of PM<sub>2.5</sub> and PM<sub>10</sub> emissions from 1<sup>st</sup> stage mote systems.

**Cotton Ginning.** Seed cotton is a perishable commodity that has no real value until the fiber and seed are separated (Wakelyn et al., 2005). Cotton must be processed or ginned at the cotton gin to separate the fiber and seed, producing 227-kg (500-lb) bales of marketable cotton fiber. Cotton ginning is considered an agricultural process and an extension of the harvest by several federal and state agencies (Wakelyn et al., 2005). Although the main function of the cotton gin is to remove the lint fiber from the seed, many other processes occur during ginning, such as cleaning, drying, and packaging the lint. Pneumatic conveying systems are the primary method of material handling in a cotton gin. As material reaches a processing point, the conveying air is separated and emitted outside the gin through a pollution control device. The amount of PM emitted by a system varies with the process and the composition of the material being processed.

Cotton ginning is a seasonal industry with the ginning season lasting from 75 to 120 days, depending on the crop size and condition. Although the general trend for U.S. cotton production has remained constant at about 17 million bales per year during the last 20 years, production from year to year often varies greatly for various reasons, including climate and market pressure. The number of active gins in the U.S. has not remained constant, but has steadily declined from 1,018 in 2000 to 682 in 2011 (NASS, 2001, 2012). Consequently, the average cotton gin production capacity across the U.S. cotton belt has increased to an approximate average of 25 bales per hour (Valco et al., 2003, 2006, 2009, 2012).

**Table 1. EPA AP-42 Appendix B.1 particle size distribution data for the battery condenser and combined lint cleaning systems equipped with cyclones on the system exhausts.**

System	% < 2.5 μm	Emission Factor kg/bale	% < 6.0 μm	Emission Factor kg/bale	% < 10 μm	Emission Factor kg/bale
Lint cleaner	1	Not Reported	20	Not Reported	54	Not Reported
Battery condenser	8	0.007	33	0.028	62	0.053

Typical cotton gin processing systems include: unloading, dryers, seed-cotton cleaners, gin stands, overflow, lint cleaners, battery condenser, bale packaging, and trash handling (Fig. 1); however, the number and type of machines and processes can vary. Each of these systems serves a unique function with the ultimate goal of ginning the cotton to produce a marketable product. Raw seed cotton harvested from the field is compacted into large units called “modules” for delivery to the gin. The unloading system removes seed cotton either mechanically or pneumatically from the module feeding system and conveys the seed cotton to the cleaning systems. Seed-cotton cleaning systems assist in drying the seed cotton and removing foreign matter prior to ginning. Ginning systems also remove foreign matter and separate the cotton fiber from seed. Lint cleaning systems further clean the cotton lint after ginning. The battery condenser and packaging systems combine lint from the lint cleaning systems and compress the lint into dense bales for efficient transport. Gin systems produce by-products or trash, such as rocks, soil, sticks, hulls, leaf material, and short or tangled immature fiber (motes), as a result of processing the seed cotton or lint. These streams of by-products must be removed from the machinery and handled by trash collection systems. These trash systems typically further process the by-products (e.g., mote cleaners) and/or consolidate the trash from the gin systems into a hopper or pile for subsequent removal.

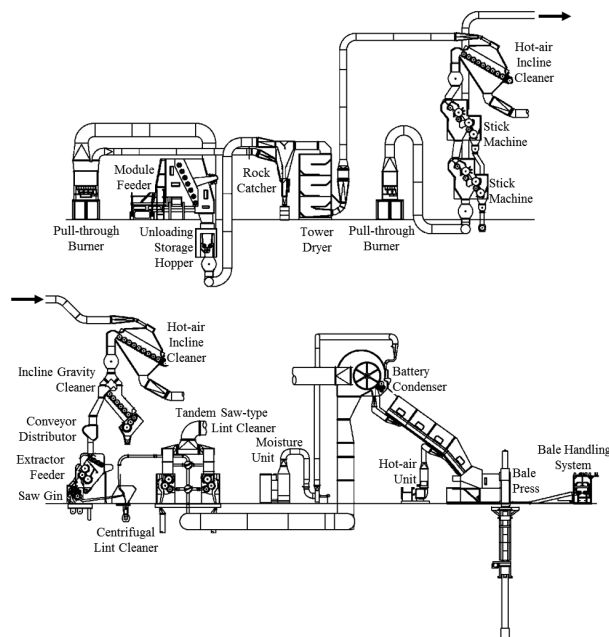


Figure 1. Typical modern cotton gin layout (Courtesy Lummus Corporation, Savannah, GA).

After the seed and lint are separated at the gin stand, the lint is cleaned by one or more stages of lint cleaners. The material removed by lint cleaners is referred to as “motes” and is handled by the mote systems (Fig. 2). Motes are pneumatically conveyed by suction away from the lint cleaners, through a centrifugal fan, to one or more particulate abatement cyclones. Depending on the gin facility, the first and second stages of lint cleaning can share a mote system, thus sharing a fan and abatement devices, or lint cleaning stages can each have mote systems that operate independently with separate fans and cyclones, denoted as 1<sup>st</sup> stage mote systems or 2<sup>nd</sup> stage mote systems. The material handled by the mote system cyclones typically includes small trash and particulate, and large amounts of lint fibers (Fig. 3).

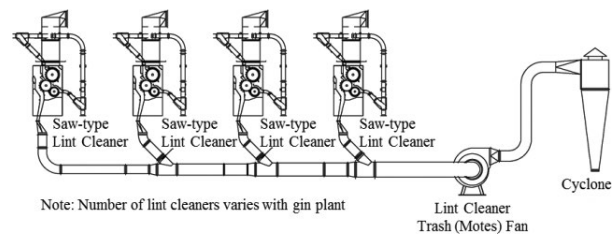


Figure 2. Typical cotton gin 1<sup>st</sup> stage mote system layout (Courtesy Lummus Corporation, Savannah, GA).

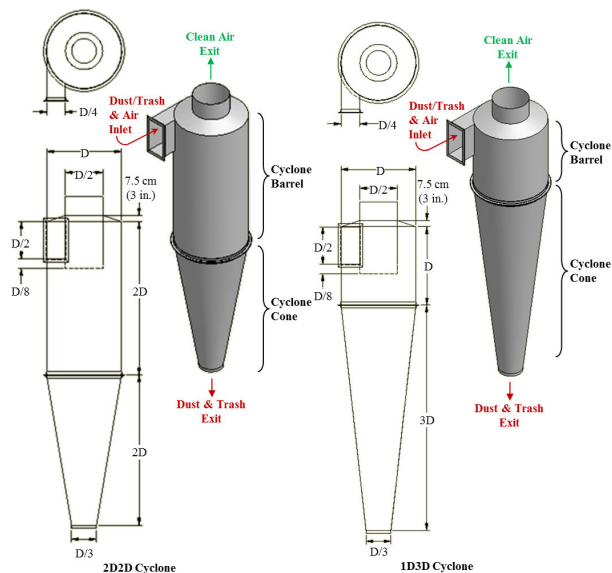


Figure 3. Photograph of typical trash captured by the 1<sup>st</sup> stage mote system cyclones.

**Cyclones.** Cyclones are the most common PM abatement devices used at cotton gins. Standard cyclone designs used at cotton ginning facilities are the 2D2D and 1D3D (Whitelock et al., 2009). The first D in the designation indicates the length of the cyclone barrel relative to the cyclone barrel diameter. The second D indicates the length of the cyclone cone relative to the cyclone barrel diameter. A standard 2D2D cyclone (Fig. 4) has an inlet height of  $D/2$  and width of  $D/4$  and design inlet velocity of  $15.2 \pm 2$



m/s ( $3000 \pm 400$  fpm). The standard 1D3D cyclone (Fig. 4) has the same inlet dimensions as either the 2D2D or the original 1D3D inlet with height of  $D$  and width  $D/8$ . Also, it has a design inlet velocity of  $16.3 \pm 2$  m/s ( $3200 \pm 400$  fpm).



**Figure 4.** 2D2D and 1D3D cyclone schematics.

**Cotton Gin Emission Factors.** EPA emission factors for cotton gins are published in EPA's Compilation of Air Pollution Emission Factors, AP-42 (EPA, 1996b). There are no 1996 EPA AP-42 emission factors for 1<sup>st</sup> stage mote systems. First stage mote systems would be similar to the mote fan listed in AP-42, but the AP-42 factor represents 1<sup>st</sup> and 2<sup>nd</sup> stage mote systems combined. The EPA AP-42 average total particulate emission factor for the mote fan was 0.13 kg (0.28 lb) per 217-kg (480-lb) equivalent bale with a range of 0.045 to 0.47 kg (0.099-1.0 lb) per bale (EPA, 1996a, b). This average and range were based on nine tests conducted in one geographical location. The EPA emission factor quality rating was D, which is the second lowest possible rating (EPA, 1996a). The AP-42 average PM<sub>10</sub> emission factor for the mote fan was 0.060 kg (0.13 lb) per 217-kg (480-lb) equivalent bale with a range of 0.023 to 0.14 kg (0.050-0.30 lb) per bale (EPA, 1996a, b). This average and range were based on six tests conducted in one geographical location, and the EPA emission factor quality rating was D. Currently there are no PM<sub>2.5</sub> emission factor data listed in the EPA AP-42 for cotton gin 1<sup>st</sup> stage mote systems.

Buser et al. (2012) discussed the plan of a large-scale project focused on developing cotton gin PM emission factors. Part of this project was focused on developing PM emission factors based on EPA-approved methodologies. Three studies focused on

1<sup>st</sup> stage mote systems evolved out of the Buser et al. (2012) project plan. Whitelock et al. (2015) reported on one study that used EPA Method 17 (CFR, 1978) to measure total particulate emission factors for the 1<sup>st</sup> stage mote systems. The system average total particulate emission factor was 0.025 kg (0.056 lb) per 227-kg (500-lb) equivalent bale with a range of 0.012 to 0.041 kg (0.026-0.089 lb) per bale. Boykin et al. (2014) reported on a second study that used EPA Method 201A (CFR, 2010) with only the PM<sub>10</sub> sizing cyclone to measure 1<sup>st</sup> stage mote system PM<sub>10</sub> and total particulate emission factors. The system average PM<sub>10</sub> and total particulate emission factors were 0.020 kg/227-kg bale (0.044 lb/500-lb bale) and 0.034 kg/bale (0.075 lb/bale), respectively. In the third study, reported by Buser et al. (2013), EPA Method 201A with both the PM<sub>10</sub> and PM<sub>2.5</sub> sizing cyclones was used to measure PM<sub>2.5</sub>, PM<sub>10</sub>, and total particulate emission factors. The average measured PM<sub>2.5</sub> emission factor was 0.0041 kg/227-kg bale (0.0090 lb/500-lb bale). The PM<sub>10</sub> and total particulate average emission factors were 0.023 kg/bale (0.051 lb/bale) and 0.032 kg/bale (0.071 lb/bale), respectively.

Particle size distribution analyses have been utilized in conjunction with total particulate sampling methods to calculate PM emissions concentration and factors for agricultural operations for more than 40 years (Wesley et al., 1972). Some examples include: cattle feedlot operations (Sweeten et al., 1998), poultry production facilities (Lacey et al., 2003), nut harvesting operations (Faulkner et al., 2009), grain handling (Boac et al., 2009), swine finishing (Barber et al., 1991), and cotton ginning (Hughes and Wakelyn, 1997). Buser and Whitelock (2007) reported cotton ginning emission concentrations based on EPA-approved PM<sub>2.5</sub>, PM<sub>10</sub>, and total particulate stack sampling methods and PSD analyses of the total particulate samples coupled with the total particulate concentrations to calculate PM<sub>2.5</sub> and PM<sub>10</sub> concentrations. The mass median diameter (MMD) of the PM in the samples ranged from 6 to 8  $\mu$ m. The study results indicated that the PSD and EPA sampler-based PM<sub>10</sub> concentrations were in good agreement, whereas the PM<sub>2.5</sub> EPA sampler concentrations ranged from 5.8 to 13.3 times the PSD-based concentrations.

The primary objective of this study was to develop PSD characteristics for the PM emitted from cotton gin 1<sup>st</sup> stage mote systems. The secondary objective was to develop PM<sub>2.5</sub> and PM<sub>10</sub> emission factors for cotton gin 1<sup>st</sup> stage mote systems equipped with cyclones on the system exhausts based on particle size

distribution analysis of total particulate samples from EPA-approved stack sampling methods.

## METHODS

Seven cotton gins were sampled across the cotton belt for the overall cotton gin sampling project described by Buser et al. (2012). Key factors for selecting specific cotton gins included: 1) facility location (geographically diverse), 2) production capacity (industry representative), 3) processing systems (typical for industry), and 4) particulate abatement technologies (properly designed and maintained 1D3D cyclones). Five of the seven gins were equipped with 1<sup>st</sup> stage mote systems that were not combined with 2<sup>nd</sup> stage mote systems. The 1<sup>st</sup> stage mote systems sampled were typical for the industry, but varied among the gins. As the lint was cleaned in three 1<sup>st</sup> stage lint cleaning systems at gin A, the trash removed from the lint was combined in the 1<sup>st</sup> stage mote system and pneumatically conveyed from the lint cleaners through a fan and exhausted through one or more cyclones. The 1<sup>st</sup> stage mote system at gin C was essentially the same, except the 1<sup>st</sup> stage mote system pulled trash from two 1<sup>st</sup> stage lint cleaning systems. The 1<sup>st</sup> stage mote systems at gins B, D, and F were also similar, but the systems at those gins pulled material from four 1<sup>st</sup> stage lint cleaning systems. Whitelock et al. (2015) provided system flow diagrams for the 1<sup>st</sup> stage mote systems that were tested.

All 1<sup>st</sup> stage mote systems sampled utilized 1D3D cyclones to control emissions (Fig. 4), but there were some cyclone design variations among the gins. Gins D and F split the system exhaust flow between two cyclones in a dual configuration (side by side as opposed to one behind another). The system airstream for gins A, B, and C was exhausted through a single cyclone. Inlets on the 1<sup>st</sup> stage mote cyclones for gins B, D, and F were 2D2D type, whereas gins A and C had inverted 1D3D inlets. Standard cones were present on 1<sup>st</sup> stage mote cyclones at all gins, except gin B, which had an expansion chamber. The cyclones tested at gins A, B, D and F had mote cyclone robber systems pulling airflow from their trash exits. This configuration helps remove lint and other trash from the cyclone that could otherwise circulate near the trash exit at the bottom of the cone for a period of time before dropping out. All of the cyclone variations outlined above, if properly designed and maintained, are recommended for controlling cotton gin emissions (Whitelock et al., 2009). Whitelock et al. (2015) provided detailed descriptions of the abatement cyclones that were tested.

**Method 17 Stack Sampling.** The samples utilized for the PSD analyses and gravimetric sample data used in developing the PSD characteristics and PSD-based emission factors were obtained from EPA Method 17 stack testing (CFR, 1978) that was conducted at the five gins with 1<sup>st</sup> stage mote systems as part of the overall cotton gin sampling project described by Buser et al. (2012). The Method 17 sampling methods and the procedures for retrieving the filter and conducting acetone wash of the sampler nozzle are described in the EPA Method 17 documentation (CFR, 1978). Further details of the project specific sampling methods, procedures, and results of the EPA Method 17 stack testing were reported by Whitelock et al. (2015).

**Laboratory Analysis.** All laboratory analyses were conducted at the USDA-ARS Air Quality Lab (AQL) in Lubbock, TX. All filters were conditioned in an environmental chamber ( $21 \pm 2^\circ\text{C}$  [ $70 \pm 3.6^\circ\text{F}$ ];  $35 \pm 5\%$  RH) for 48 h prior to gravimetric analyses. Filters were weighed in the environmental chamber on a Mettler MX-5 microbalance (Mettler-Toledo Inc., Columbus, OH; 1  $\mu\text{g}$  readability and 0.9  $\mu\text{g}$  repeatability) after being passed through an antistatic device. The MX-5 microbalance was leveled on a marble table and housed inside an acrylic box to minimize the effects of air currents and vibrations. To reduce recording errors, weights were digitally transferred from the microbalance directly to a spreadsheet. Technicians wore latex gloves and a particulate respirator mask to avoid contaminating the filter or sample. AQL procedures required that each sample be weighed three times. If the standard deviation of the weights for a given sample exceeded 10  $\mu\text{g}$ , the sample was reweighed. Gravimetric procedures for the acetone wash tubs were the same as those used for filters.

In addition to gravimetric analyses, each sample was visually inspected for unusual characteristics, such as cotton lint content or extraneous material. Digital pictures were taken of all filters and washes for documentation purposes. After the laboratory analyses were completed all stack sampling, cotton gin production, and laboratory data were merged.

**Particle Size Analysis.** A Beckman Coulter LS230 laser diffraction system (Beckman Coulter Inc., Miami, FL) with software version 3.29 was used to perform the particle size analyses on the filter and wash samples. The instrument sizes particles with diameters ranging from 0.4 to 2000  $\mu\text{m}$ . For this project, the LS230 fluid module was used with a 5% lithium chloride/methanol suspension fluid mixture. Approximately 10-L batches of the suspension fluid

were prepared and stored in a self-contained, recirculating, filtration system equipped with 0.2 μm filters to keep the fluid well mixed and free of larger particles. Prior to each test run a background particle check was performed on the fluid to help minimize particulate contamination from non-sample sources. The process of analyzing the samples included the following steps:

1. pour approximately 40 mL of clean suspension fluid into a clean 100-mL beaker;
2. transfer a particulate sample to the 100-mL beaker with clean suspension fluid,
  - a. for 47-mm filter media, remove the filter from the Petri dish with tweezers and place the filter in the 100-mL beaker with the suspension fluid,
  - b. for the wash samples contained in a sample tub, use a small amount of the suspension fluid and a sterile foam swab to transfer the sample from the tub to the 100-mL beaker;
3. place the 100-mL beaker in an ultrasonic bath for 5 min to disperse the PM sample in the fluid;
4. using a sterile pipette, gradually introduce the PM and suspension fluid mixture into clean suspension fluid that is being monitored by the LS230 until an obscuration level of 10% is reached;
5. activate the LS230 system to measure the diffraction patterns and calculate the PSD;
6. repeat step 5 a total of three times and average the results; and
7. drain and flush/clean the LS230 system.

Optical models for calculating laser diffraction-based PSDs require input of a refractive index for the suspension fluid and real and imaginary refractive indices for the sample. A refractive index of 1.326 for methanol was used for the suspension fluid (Beckman Coulter, 2011). Hughs et al. (1997) showed that particulate from cyclone exhausts was about 34% ash or fine soil particulate with the balance made up of water and organic material (e.g., cellulose, lignin, protein). Real and imaginary refractive index values for common soil constituents—quartz, clay minerals, silica and feldspars—are 1.56 and 0.01, respectively (Buurman et al., 2001). These indices were used in the optical model used in calculating the PSD for the cyclone particulate samples. Wang-Li et al. (2013) and Buser (2004) provided additional details on the PSD methodology.

The LS230 PSD results are in the form of particle volume versus equivalent spherical diameter.

The PSD results were converted to particle volume versus AED using the following equation:

$$d_a = d_p \left( \frac{\rho_p}{\kappa \rho_w} \right)^{1/2}$$

where  $\rho_w$  is the density of water with a value of 1 g/cm<sup>3</sup>,  $\rho_p$  is the particle density, and  $\kappa$  is the dynamic shape factor. The dynamic shape factor was determined to be 1.4 based on Hinds (1982) factors for quartz and sand dust. The particle density, assumed to be constant for the Method 17 filter and wash samples evaluated in this study, was determined in an earlier study to be 2.65 g/cm<sup>3</sup> (M. Buser, unpublished data, 2013). This earlier study used a helium displacement AccuPyc 1330 Pycnometer (Micromeritics, Norcross, GA) to determine the particle density of cotton gin waste that passed through a No. 200 sieve (particles that pass through a 74-μm sieve opening). The study was based on three random samples collected at 43 different cotton gins.

Results obtained from each average adjusted PSD included: MMD, mass fraction of PM with diameter less than or equal to 10 μm (PM<sub>10</sub>), mass fraction of PM with diameter less than or equal to 6 μm (PM<sub>6</sub>), and mass fraction of PM with diameter less than or equal to 2.5 μm (PM<sub>2.5</sub>). This information was coupled with the corresponding Method 17 sample mass to calculate the PM<sub>10</sub>, PM<sub>6</sub>, and PM<sub>2.5</sub> emission factors using the following equation:

$$EF_i = EF_{tot} \left( \left( \frac{M_F}{M_F + M_W} \right) w_{Fi} + \left( \frac{M_W}{M_F + M_W} \right) w_{Wi} \right)$$

where  $EF_i$  = emission factor for particle in the size range  $i$ ;

$EF_{tot}$  = total particulate emission factor obtained from total particulate tests (Whitelock et al., 2015);

$M_F$  = total mass of particulate on filter;

$M_W$  = total mass of particulate in nozzle wash;

$w_{Fi}$  = mass fraction of particles on the filter in the size range  $i$ ; and

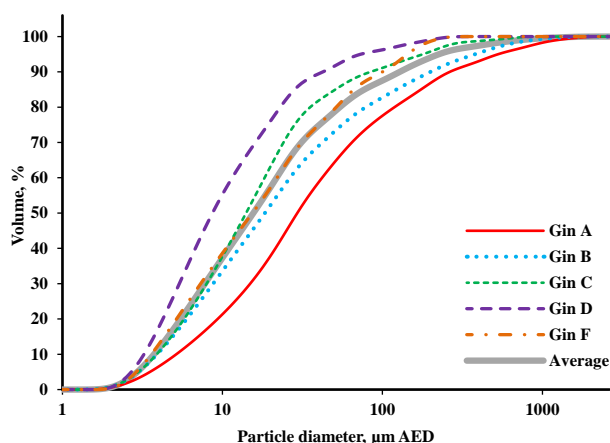
$w_{Wi}$  = mass fraction of particles in the nozzle wash in the size range  $i$ .

The 1<sup>st</sup> stage mote systems sampled were typical for the industry. The system average ginning rate was 27.9 bales/h and the test average ginning rate at each gin ranged from 21.2 to 35.7 bales/h (based on 227-

kg [500-lb] equivalent bales). The capacity of gins sampled was representative of the industry average, approximately 25 bales/h. The 1D3D cyclones were all operated with inlet velocities within design criteria,  $16.3 \pm 2$  m/s ( $3200 \pm 400$  fpm). There are criteria specified in EPA Method 17 for test runs to be valid for total particulate measurements (CFR, 1978). Isokinetic sampling must fall within EPA-defined range of  $100 \pm 10\%$ . All tests met the isokinetic criteria. The stack gas temperatures ranged from 23 to 41°C (73-106°F) and moisture content ranged from 0.2 to 2.7%. The individual systems and cyclone design variations were discussed by Whitelock et al. (2015).

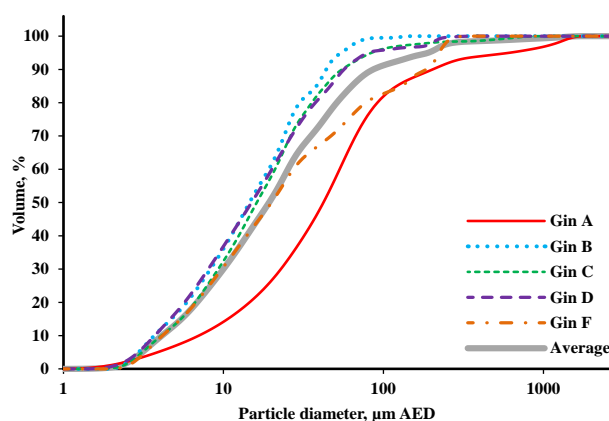
## RESULTS

The PSD characteristics and mass of the PM captured on the filters are shown in Table 2. The mass of the PM captured on the filter accounted for 65 to 94% of the total PM (filter and wash) collected from the individual test runs. The system average MMD for particulate on the filters was  $15.6 \mu\text{m AED}$ . Test averages ranged from 8.7 to  $29.4 \mu\text{m AED}$ . The test and system averages are based on averaging PSDs and not averaging individual test results. The mass fraction of  $\text{PM}_{2.5}$ ,  $\text{PM}_6$ , and  $\text{PM}_{10}$  ranged from 1.77 to 3.49%, 12.3 to 34.7%, and 21.3 to 55.2%, respectively. Filter PM PSDs for the five gins and the system average are shown in Fig. 5. In general, the PSD curves for the PM captured on the filters for gins had similar shapes. The shift to the right and wider particle diameter range illustrates the larger MMD of the gin A distribution, whereas the PSD for gin D exhibits characteristics of much smaller MMD.



**Figure 5.** Gin average cumulative particle size distributions for the PM captured on a EPA-Method 17 filter from the 1<sup>st</sup> stage mote systems.

The PSD characteristics and mass of the PM captured in the washes are shown in Table 3. The mass of the PM captured in the sampler nozzle and retrieved in the wash accounted for 6 to 35% of the total PM (filter and wash) collected from the individual test runs. The system average MMD was  $19.3 \mu\text{m AED}$ . Test average MMDs ranged from 14.6 to  $42.2 \mu\text{m AED}$ . The mass fraction of  $\text{PM}_{2.5}$ ,  $\text{PM}_6$ , and  $\text{PM}_{10}$  ranged from 0.97 to 2.56%, 8.5 to 21.3%, and 14.2 to 36.8%, respectively. PSDs for the PM captured in the nozzle for the five gins and the system average are shown in Fig. 6. In general, the PSD curves for the PM captured in the washes had similar shapes. Like the filter PSDs, the gin A PSD illustrates the larger MMD.



**Figure 6.** Gin average cumulative particle size distributions for the PM captured in the EPA-Method 17 sampler nozzle wash from the 1<sup>st</sup> stage mote systems.

The combined PSD characteristics for the PM captured on the filter and PM captured in the wash are shown in Table 4. The 1<sup>st</sup> stage mote system average combined filter and wash PSD MMD was  $16.4 \mu\text{m AED}$  (9.3 to  $31.1 \mu\text{m}$  test average range). There were no particles less than  $0.5 \mu\text{m}$  and less than 0.01% of the particles had a diameter of  $1 \mu\text{m}$  or smaller. The combined filter and wash  $\text{PM}_{2.5}$ ,  $\text{PM}_6$ , and  $\text{PM}_{10}$  mass fractions ranged from 1.84 to 3.37%, 11.9 to 32.9%, and 20.5 to 52.7%, respectively. Combined PM PSDs for the five gins and the system average are shown in Fig. 7. These combined PSDs were similar among gins and were more consistent with the filter PSDs than the wash PSDs. This was expected because the majority of the PM mass was captured on the filter as compared to the nozzle wash.

Table 2. EPA Method 17 filter particle size distribution data for the 1<sup>st</sup> stage mote system.

Gin	Test Run	Mass Median Diameter µm AED	PM <sub>2.5</sub> %	PM <sub>6</sub> %	PM <sub>10</sub> %	Sample Total mg
A	1 <sup>z</sup>	–	–	–	–	–
	2	30.3	1.66	12.2	21.2	92.33
	3	28.5	1.88	12.4	21.5	104.22
	Test Average (n = 2) <sup>y</sup>	29.4	1.77	12.3	21.3	
B	1	15.3	2.77	23.8	38.5	18.84
	2	18.1	2.99	19.4	33.2	11.68
	3	22.3	2.02	16.9	29.1	23.27
	Test Average (n = 3) <sup>y</sup>	18.4	2.59	20.1	33.6	
C	1	13.1	3.04	23.2	40.2	12.69
	2	15.6	2.30	18.6	34.1	22.39
	3	13.6	2.60	21.7	38.8	10.88
	Test Average (n = 3) <sup>y</sup>	14.1	2.65	21.2	37.7	
D	1	8.8	3.33	34.0	54.5	26.31
	2	8.8	3.55	34.7	54.7	14.55
	3	8.4	3.58	35.5	56.5	20.96
	Test Average (n = 3) <sup>y</sup>	8.7	3.49	34.7	55.2	
F	1	18.2	2.42	21.2	34.4	7.36
	2	15.2	2.20	25.1	39.1	7.15
	3	13.2	2.85	26.6	42.3	9.82
	Test Average (n = 3) <sup>y</sup>	15.4	2.49	24.3	38.6	
System Average (n = 5) <sup>y</sup>		15.6	2.60	22.5	37.3	

<sup>z</sup> Insufficient sample for particle size distribution

<sup>y</sup> Based on averaged particle size distributions

Table 3. EPA Method 17 nozzle wash particle size distribution data for the 1<sup>st</sup> stage mote system.

Gin	Test Run	Mass Median Diameter µm AED	PM <sub>2.5</sub> %	PM <sub>6</sub> %	PM <sub>10</sub> %	Sample Total mg
A	1	36.4	2.09	8.6	14.3	6.81
	2	36.2	3.06	10.9	18.0	14.57
	3	61.5	1.46	6.0	10.2	12.80
	Test Average (n = 3) <sup>z</sup>	42.2	2.20	8.5	14.2	
B	1	15.9	1.92	19.0	33.8	2.42
	2	12.3	2.32	24.2	41.7	1.70
	3	16.1	1.73	18.2	32.7	1.43
	Test Average (n = 3) <sup>z</sup>	14.6	1.99	20.5	36.1	
C	1	30.9	1.68	10.3	18.8	6.74
	2	13.7	1.45	18.4	36.1	1.71
	3	12.1	1.75	23.2	42.3	1.33
	Test Average (n = 3) <sup>z</sup>	16.2	1.63	17.3	32.4	
D	1	11.9	4.31	25.1	43.1	1.96
	2	17.6	2.73	19.6	33.1	4.63
	3	16.9	0.65	19.2	34.2	1.76
	Test Average (n = 3) <sup>z</sup>	15.0	2.56	21.3	36.8	
F	1	16.7	1.18	19.2	33.4	1.48
	2	23.2	1.15	16.3	28.5	1.87
	3	22.4	0.57	16.0	30.1	1.05
	Test Average (n = 3) <sup>z</sup>	20.0	0.97	17.2	30.6	
System Average (n = 5) <sup>z</sup>		19.3	1.87	16.9	30.0	

<sup>z</sup> Based on averaged particle size distributions



Table 4. EPA Method 17 combined filter and wash particle size distribution data for the 1<sup>st</sup> stage mote system.

Gin	Test Run	Mass Median Diameter µm AED	PM <sub>2.5</sub> %	PM <sub>6</sub> %	PM <sub>10</sub> %
A	1 <sup>z</sup>	–	–	–	–
	2	31.1	1.85	12.1	20.7
	3	31.0	1.84	11.7	20.2
	Test Average (n = 2) <sup>y</sup>	31.1	1.84	11.9	20.5
B	1	15.4	2.67	23.3	37.9
	2	17.0	2.91	20.0	34.3
	3	21.7	2.00	17.0	29.3
	Test Average (n = 3) <sup>y</sup>	17.9	2.53	20.1	33.8
C	1	17.2	2.57	18.7	32.8
	2	15.4	2.24	18.6	34.3
	3	13.4	2.51	21.9	39.1
	Test Average (n = 3) <sup>y</sup>	15.2	2.44	19.7	35.4
D	1	9.0	3.40	33.3	53.8
	2	10.2	3.35	31.1	49.5
	3	8.8	3.35	34.3	54.8
	Test Average (n = 3) <sup>y</sup>	9.3	3.37	32.9	52.7
F	1	17.9	2.21	20.8	34.2
	2	16.6	1.98	23.3	36.9
	3	13.8	2.63	25.5	41.1
	Test Average (n = 3) <sup>y</sup>	16.0	2.28	23.2	37.4
System Average (n = 5) <sup>y</sup>		16.4	2.49	21.6	36.0

<sup>z</sup> Insufficient sample for particle size distribution

<sup>y</sup> Based on averaged particle size distributions

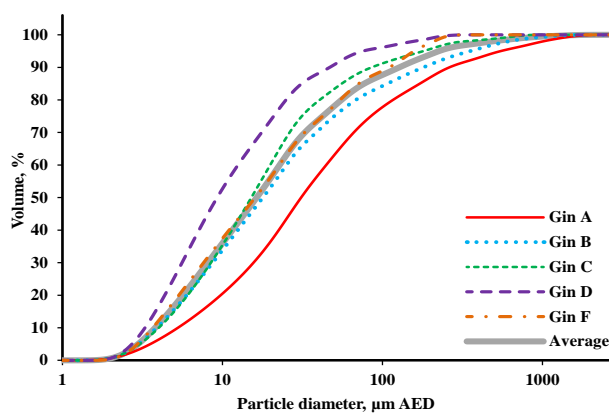


Figure 7. Gin average cumulative particle size distributions for the EPA-Method 17 combined filter and wash samples from the 1<sup>st</sup> stage mote systems.

The PSD-based emission factors for the 1<sup>st</sup> stage mote systems are shown in Table 5. The system average PM<sub>2.5</sub> emission factor was 0.00063 kg/227-kg (500-lb) bale (0.0014 lb/bale). PM<sub>2.5</sub> emission factors ranged from 0.00021 to 0.0017 kg (0.00047-0.0038 lb) per bale. The 1<sup>st</sup> stage mote system average PM<sub>6</sub> emission factor was 0.0054

kg/bale (0.012 lb/bale). The PM<sub>10</sub> emission factors ranged from 0.0019 to 0.017 kg/bale (0.0041-0.038 lb/bale). The 1<sup>st</sup> stage mote system average PM<sub>10</sub> emission factor was 0.0091 kg/bale (0.020 lb/bale) and ranged from 0.0033 to 0.028 kg (0.0073-0.061 lb) per bale. The ratios of PM<sub>2.5</sub> to total particulate, PM<sub>6</sub> to total particulate, and PM<sub>10</sub> to total particulate, based on the system averages, were 2.49, 21.6, and 36%, respectively.

The PSD-based 1<sup>st</sup> stage mote system PM<sub>2.5</sub> emission factor was approximately 15% of the PM<sub>2.5</sub> emission factor reported by Buser et al. (2013) and measured using EPA Method 201A, 0.0041 kg (0.0090 lb) per 227-kg (500-lb) bale. The PSD-based 1<sup>st</sup> stage mote system PM<sub>10</sub> emission factor was 15% of the EPA AP-42 published value for the mote fan, 0.060 kg (0.13 lb) per bale (EPA, 1996a), which is an equivalent system to the combined 1<sup>st</sup> and 2<sup>nd</sup> stage mote system. Also, the PSD-based system PM<sub>10</sub> emission factor was 45% of the Method 201A (PM<sub>10</sub> sizing cyclone only) PM<sub>10</sub> emission factor reported by Boykin et al. (2014), 0.020 kg (0.044 lb) per bale. The PSD-based PM<sub>10</sub> emission

factor was 40% of the Method 201A (PM<sub>10</sub> and PM<sub>2.5</sub> sizing cyclones) PM<sub>10</sub> emission factor reported by Buser et al. (2013), 0.023 kg (0.051 lb) per bale. The differences among the methods could be attributed to several sources. First, due to constraints in the EPA methods, the three studies utilizing Method 17 for total particulate sampling and PSD analyses, Method 201A for PM<sub>10</sub> sampling, and Method 201A for PM<sub>2.5</sub> and PM<sub>10</sub> sampling could not be conducted simultaneously. Combined with the fact that emissions from cotton ginning can vary with the condition of incoming cotton, PM concentrations measured among the three studies could have varied. Second, for reasons described by Buser (2007a, b, c) and documented by Buser and Whitelock (2007), some larger particles could penetrate the Method 201A sampler PM<sub>10</sub> or PM<sub>2.5</sub> sizing cyclones and collect on the filter. Finally, cotton fibers have a cross-sectional diameter much larger than 10 μm and are difficult to scrub out of air streams. These fibers could cycle in the sizing cyclones and pass through to deposit on the filters. This behavior was observed

during some of the Method 201A testing where cotton fibers were found in Method 201A sampler washes and on filters (Fig. 8). Currently there are no EPA-approved guidelines to adjust Method 201A PM<sub>10</sub> or PM<sub>2.5</sub> concentration measurements to account for these fibers.

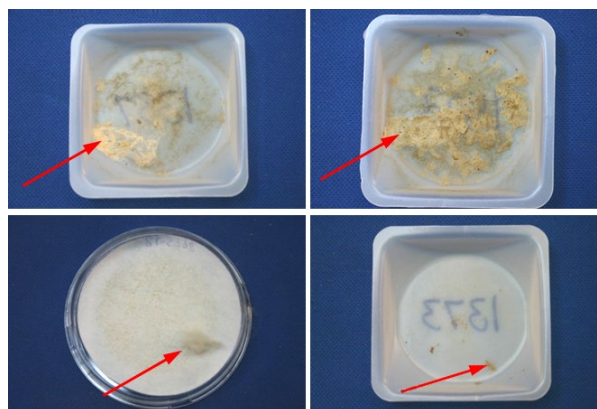


Figure 8. Example EPA Method 201A filter and sampler head acetone washes with lint (indicated by arrows) in the washes and on the filter. Clockwise from top left: > 10 μm wash, 10 to 2.5 μm wash, ≤ 2.5 μm wash, and filter.

Table 5. EPA Method 17 total particulate and particle size distribution-based PM<sub>2.5</sub>, PM<sub>6</sub>, and PM<sub>10</sub> emission factor data for the 1<sup>st</sup> stage mote system.

Gin	Test Run	Total <sup>x</sup>		PM <sub>2.5</sub> <sup>w</sup>		PM <sub>6</sub> <sup>w</sup>		PM <sub>10</sub> <sup>w</sup>	
		kg/bale <sup>z</sup>	lb/bale <sup>z</sup>	kg/bale <sup>z</sup>	lb/bale <sup>z</sup>	kg/bale <sup>z</sup>	lb/bale <sup>z</sup>	kg/bale <sup>z</sup>	lb/bale <sup>z</sup>
A	1 <sup>y</sup>	0.039	0.086	–	–	–	–	–	–
	2	0.040	0.089	0.00075	0.0016	0.0049	0.011	0.0084	0.018
	3	0.042	0.093	0.00077	0.0017	0.0049	0.011	0.0085	0.019
B	1	0.017	0.038	0.00046	0.0010	0.0040	0.0087	0.0065	0.014
	2	0.023	0.050	0.00066	0.0015	0.0046	0.010	0.0078	0.017
	3	0.020	0.045	0.00041	0.00090	0.0035	0.0076	0.0060	0.013
C	1	0.014	0.032	0.00037	0.00081	0.0027	0.0059	0.0047	0.010
	2	0.018	0.041	0.00041	0.00091	0.0034	0.0075	0.0063	0.014
	3	0.0085	0.019	0.00021	0.00047	0.0019	0.0041	0.0033	0.0073
D	1	0.051	0.113	0.0017	0.0038	0.017	0.038	0.028	0.061
	2	0.032	0.070	0.0011	0.0023	0.010	0.022	0.016	0.035
	3	0.038	0.083	0.0013	0.0028	0.013	0.028	0.021	0.045
F	1	0.011	0.025	0.00025	0.00055	0.0024	0.0052	0.0039	0.0085
	2	0.011	0.024	0.00022	0.00048	0.0026	0.0057	0.0041	0.0089
	3	0.012	0.028	0.00033	0.00072	0.0032	0.0070	0.0051	0.011
System Average		0.025	0.056	0.00063	0.0014	0.0054	0.012	0.0091	0.020

<sup>z</sup> 227-kg (500-lb) equivalent bales

<sup>y</sup> Insufficient sample for particle size distribution

<sup>x</sup> Taken from Whitelock et al. (2015)

<sup>w</sup> Factors are the product of the corresponding PM percentage from Table 4 and the total particulate emission factor

## SUMMARY

Cotton gins across the U.S. cotton belt were sampled using EPA-approved methods to fill the data gap that exists for PM<sub>2.5</sub> cotton gin emissions data and to collect additional data to improve the EPA AP-42 total and PM<sub>10</sub> emission factor quality ratings for cotton gins. Samples were further analyzed to characterize the PSD of the particulate measured. Five selected cotton gins had 1<sup>st</sup> stage mote systems that used pneumatic conveyance and had exhaust airstreams that were not combined with another system. All tested systems were similar in design and typical of the ginning industry and were equipped with 1D3D cyclones for emissions control. In terms of capacity, the five gins were typical of the industry, averaging 27.9 bales/h during testing. The average PSD-based 1<sup>st</sup> stage mote system PM<sub>2.5</sub>, PM<sub>6</sub>, and PM<sub>10</sub> emission factors from the five gins tested (14 total test runs) were 0.00063 kg/227-kg bale (0.0014 lb/500-lb bale), 0.0054 kg/bale (0.012 lb/bale), and 0.0091 kg/bale (0.020 lb/bale), respectively. The PSDs were characterized by an average MMD of 16.4 μm AED. Based on system average emission factors, the ratio of PM<sub>2.5</sub> to total particulate was 2.49%, PM<sub>6</sub> to total particulate was 21.6%, and PM<sub>10</sub> to total particulate was 36.0%. PSD-based system average PM<sub>2.5</sub> and PM<sub>10</sub> emission factors were 15% and 45% of those measured for the overall cotton gin sampling project utilizing EPA-approved methods. The PSD-based PM<sub>10</sub> emission factor was 15% of that currently published in EPA AP-42 for the mote fan, which is equivalent to a combined 1<sup>st</sup> and 2<sup>nd</sup> stage mote system. Currently there are no 1<sup>st</sup> stage mote system emission factors published EPA AP-42.

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